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Abstract
Arsenic (As) exposure is a global public health problem because of its association with various cancers and numerous other pathological effects, and millions of people worldwide are exposed to As on a regular basis mainly through drinking water. Increasing lines of evidence indicate that As may adversely affect the antioxidant defense system, but its specific mechanism to abrupt the antioxidant defense system are poorly understood. Therefore, we conducted a literature search of As and its oxidative stress-related effects associated with As exposure and summarized the known oxidative disorders of As in humans and laboratory animals. Overall, the review indicates that chronic exposure to As has the potential to impair the antioxidant system which could lead to increased risk of disorders and chronic diseases, including various cancers. Further investigation, particularly in humans, is needed to better understand the relationship between As exposure and the development of disease as well as the proper mechanism.

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Oxidative stress

Introduction
Most of the world has been polluted due to the contaminated toxic effects of Arsenic (As). It is mobile in nature. From the rocks due to weathering it has been released to the environment and dissolve to the environment (Flora et al., 2005; Mandal and Suzuki, 2002). Chronic exposure to inorganic arsenic can lead deterioration of several organs in the body after resulting of cancer (Aposhian et al., 2003). With different chemical forms it has different types of toxicity depending on its doses also. Mostly, inorganic and organic forms are abundant and among them inorganic forms has most deleterious effects (Geiszinger et al., 2002; Fattorini and Regoli, 2004). Among the other arsenic species, moderately toxicity has been showed by trimethyl-arsine oxide (TMAO) and tetra- methyl-arsonium (TETRA). Meanwhile, arsenobetaine (AsB), arsenocholine (AsC) and other arsenosugars (AsS) show no toxicity (Fattorini et al., 2006). Arsenic caused the imbalance between pro-oxidant and anti-oxidant which ultimately showed the oxidative stress (Shila et al., 2005). Moreover, due to presence of differences in the chemical form, it has has many differences between arsenite and arsenate effects (Tseng, 2004). Because of its similar biochemical properties to phosphate, arsenate integrated with the phosphorylation reactions (Fattorini and Regoli, 2004) and has the biochemical properties to sulfydryl groups, which can induce structural modification in proteins, leading to the inactivation of many enzymes (Akter et al., 2005). Inorganic arsenic has been classified as a known human carcinogen (IRIS,
An overview of arsenic (As)

Sources of arsenic in the environment: Arsenic (As) is a naturally occurring and wide spread pollutant in various regions of the world with its both metalloid and non metalloid properties (Flora et al., 2005). Arsenic and its compounds have showed the mobile nature in environment. Arsenic sulfides to arsenic trioxide have been converted due to the weathering process of rocks. These compounds then entered in the environment through dust, rain, river water or groundwater, etc. (Mandal and Suzuki, 2002). Arsenic can be found in rock, soil, water, air and the earth biosphere. There are many forms or species of arsenic and these can be broadly categorized as inorganic or organic (EFSA, 2009). Arsenic is also found in surface rocks and minerals, where it may be combined with sulfur and/or metals, including manganese, iron, cobalt, tin, silver, and nickel often as a complex oxide (Waugh, 1982).

Chemical characteristics of As

Arsenic: atomic number 33 is ranking 20th in natural abundance, comprising about 0.00005% of the earth’s crust, 14th in the seawater, and 12th in the human body (Mandal and Suzuki, 2002). It’s concentration in most rocks ranges from 0.5 to 2.5 mg/kg, though higher concentrations are found in finer grained argillaceous sediments and phosphorites (Mandal and Suzuki, 2002). It is a silver-grey brittle crystalline solid with atomic weight 74.9; specific gravity 5.73, melting point 817°C (at 28 atm), boiling point 613°C and vapor pressure 1mm Hg at 372°C. It was isolated in 1250 a.d. by Albertus Magnus (Mandal and Suzuki, 2002).

Types of As

Arsenic exists in the –3, 0, +3 and +5 oxidation states (Smedley and Kinniburgh, 2002). There are many forms, or species, of arsenic and these can be broadly categorized as inorganic or organic (Aronson, 1994). Two forms are common in natural waters: arsenite (AsO\(_3^{-3}\)) and arsenate (AsO\(_4^{3-}\)), referred to as arsenic (III) and arsenic (V). Pentavalent (+5) or arsenite species are AsO\(_4^{3-}\), HAsO\(_4^{2-}\), H\(_2\)AsO\(_4^{-}\) while trivalent (+3) arsenates include As(III)\(_2\), As(III)\(_3\)\(_2\), AsO\(_2\)\(_3\)\(_2\) and AsO\(_3\)\(_3\). Pentavalent species predominate and are stable in oxygen rich aerobic environments.

Trivalent arsenites predominate in moderately reducing anaerobic environments such as groundwater (EFSA, 2009). While methylated forms (methylarsonate, MMA and dimethyl arsinite, DMA) are considered only moderately toxic (Fattorini and Regoli, 2004). Other arsenic species, like trimethyl-arsineoxide (TMAO) and tetra- methyl-arsonium (TETRA) are also considered moderately toxic, whereas arsenobetaine (AsB), arsenocholine (AsC) and other arsano sugars (AsS) show no toxicity (Fattorini et al., 2006). Environmental forms include arsenious acids (H\(_2\)AsO\(_3\)), H\(_3\)AsO\(_4\)), arsenic acids (H\(_5\)AsO\(_4\)\(_3\)), arsenates, arsenites, arsenates, methylarsenic acid, dimethylarsinic acid, arsenic, etc. Arsenic (III) is hard acidic in nature which and preferably chelates with oxides and nitrogen. Whereas, arsenic which has the valences with (V) shows soft acidic nature and has the great tendency to form complexes with sulfides (Bodek, 1998).

Metabolism of As in animal body

The main inorganic forms of arsenic relevant for human exposures are pentavalent arsenic (also called arsenate, As (V), or As\(^{+5}\)) and trivalent arsenic (also called arsenite, As (III), or As\(^{+3}\)). These inorganic species undergoes a series of reduction and oxidative/methylation steps in human liver and other tissues to form tri- and pentavalent methylated metabolites of methylarsonite [MA(III)], methylarsonate [MA(V)], dimethylarsinite [DMA(III)], and dimethylarsinate [DMA(V)].

Some mammalian species also produce trimethylated metabolites, trimethylarsine oxide [TMA (V)O] and, possibly, the volatile trimethylarsine [TMA(V)]. Fish and other seafood are the major sources of exposure to organic arsenic, in the form of organobetaine, arsenosugars, and arsenolipids. The distinction between inorganic and organic forms is important because it is generally accepted that the organic species are excreted more quickly from the body and generally considered less toxic, with a relative rank order of As(III) > As(V) >>
MA(V), DMA(V) >> arsenobetaine. However, the methylated trivalent metabolites, MA(III) and DMA(III), are significantly more toxic than their pentavalent counterpart and either As(III) or As(V) (Akter et al., 2005). Inorganic arsenic is metabolized by a sequential process and described by Vahter (1994) as below (Fig. 1):

**Fig. 1:** Inorganic arsenic metabolism by a sequential process (Source: Vahter, 1994).

### Biotransformation of As

In drinking water, As is normally found as As\(^{V}\). After the consumption by humans and other organisms it rapidly undergoes metabolic conversion. This metabolic conversion called as biotransformation (Sakurai et al., 2005). Through this process, at first, the arsenate is converted into arsenite and then subsequently transformed into mono-, di-, and trimethylated products (Thomas et al., 2004). These reactions show that As methylation is associated with reduction of As\(^{V}\) to As\(^{III}\) (Thomas et al., 2004). As a reducing agent, antioxidant glutathione (GSH) plays an important role in arsenic biotransformation. GSH convert As to its trivalency and formed arsenotriliglutathione [As\(^{III}\) (GS)\(_3\)]. In this complex, As\(^{III}\) is joint to the thiol groups of the cysteinyl residues of three GSH molecules (Kobayashi et al., 2005; Thomas et al., 2001). Mainly, reduction reaction of arsenate to arsenite catalyzed by anarsenate reductase enzyme. This process requires the presence of inosineandathiol compound (Aposhian et al., 2004).

Aarsenite is methylated by arsenate methyltransferase in mammals (Akter et al., 2005). One of the *in-vitro* study on rat liver, exposed with arsenite and methylarsonous diiodide (CH\(_3\)As\(^{IIII}\)\(_2\)) demonstrated that arsenite was the preferred substrate for the ethylation reaction. The reduction conversion reaction of arsenite to methylated metabolites being faster than for arsenate. For the arsenic methylation, a donor of methyl groups must be present at the time of reduction conversion reaction. Studies, both on, *in vitro* and *in vivo* with the mammal models have been identified that S-adenosylmethionine (AdoMet) act as the methyl group donor (Thomas et al., 2004; Thomas et al., 2001). The enzyme methyl arsonate reductase catalyzes the reduction of monomethylarsonate (MMA\(^V\)), dimethylarsonate (DMA\(^V\)) and arsenate (As\(^V\)) to monomethyl arsonous acid (MMA\(^{III}\)), dimethyl arsonous acid (DMA\(^{III}\)) and arsenite (As\(^{III}\)), respectively. The activity of this enzyme is the rate-limiting step for inorganic arsenic methylation. Methyl arsonate reductase has an absolute requirement for GSH, being recognized as the omegaiso form of the enzyme glutathione-S-
transferase (GST) (Zakharyan et al., 1999; Aposhian et al., 2004). GST families are multifunctional molecules. This plays a key role in cellular detoxification. This family of enzyme is the rate-limiting enzyme for biotransformation of inorganic arsenic (Townsend and Tew, 2003; Sampayo-Reyes and Zakharyan, 2006).

**Effect of As throughout globe**

Worldwide billions of people are obligated to consume the drinking water with higher arsenic level than the prescribed level of World Health Organization (WHO). Now a day, it is the worldwide problem of As contaminated natural water consumption. Most of the countries like USA, China, Chile, Bangladesh, Taiwan, Mexico, Argentina, Poland, Canada, Hungary, New Zealand, Japan and India has been reported with As contamination. In the area of Indo–Bangladesh region, about 70 million people are suffering from As contaminated drinking water consumption (Karim, 2000; Hassan et al., 2003).

According to Peters et al. (2000) in New Hampshire, USA, inorganic arsenic was present in about 95% sample of drinking water. In these sample, the level of As lied within the range of 0.01 μg/l to 180 μg/l. In Argentina, the concentration of arsenic in groundwater ranged from 100 μg/l to 2000 μg/l. In this country about 200,000 people are consume the contaminated water (British Geological Survey, 2001). In case of Romania and Hungary, about 4,00,000 population used drinking water above minimum concentration level (MCL). The range of arsenic concentration was 2 to 176 μg/l (WHO, 2003). After the study of Smith et al. (2000), in Chile, it was found that the range of arsenic level in drinking water was 750 to 800 μg/l.

In China, the concentration of arsenic in well water in the affected areas was 50 μg/l to 2000 μg/l and about 2 million people in the affected area use the drinking well water containing arsenite more than common standard, which cause Raynaud’s disease in the population (Xia and Liu, 2004). In Taiwan, the arsenic concentration in well water used for drinking purpose were 10-1800 μg/l and a peripheral vesicular disease called “Black foot disease” is a common disease among the living population due to arsenicism (Lamm et al., 2006).

Report of British Geological Survey (2001) informed that in Bangladesh, more than 70-80 million people are at a risk of drinking contaminated water. The drinking water arsenic levels were ranged from non-detectable to 4700 μg/l. West Bengal (India) and Bangladesh are the worst affected areas in the world from arsenicism. The standard of most developing countries is 50 μg/l, which is several times higher than the MCL and more hazardous to the population.

The study of Acharya (2002) stated that in West Bengal (India), the arsenic concentration in drinking water is about 60 to 3700 μg/l. About 40 million people are affected from this reason. Another study of Chakraborti et al. (2003), demonstrated that in middle Ganga plain, Bihar, 206 tube wells (95% of total) were analyzed for arsenic content and showed that 56.8% tube wells have exceeded arsenic concentration of 50 μg/l and 19.9% have more than 300 μg/l. The maximum permissible limits for drinking water in different countries are given in the Table 1.

Table 1. The maximum permissible limits for drinking water in different countries.

<table>
<thead>
<tr>
<th>Sl. No.</th>
<th>Sources</th>
<th>Name of the country</th>
<th>Maximum permissible limit (mg/l)</th>
</tr>
</thead>
<tbody>
<tr>
<td>1.</td>
<td>Bhattacharya et al. (2006)</td>
<td>Argentina</td>
<td>50</td>
</tr>
<tr>
<td>2.</td>
<td>Kinniburgh and Smedley (2001)</td>
<td>Bangladesh</td>
<td>50</td>
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<tr>
<td>5.</td>
<td>Dhar et al. (1997)</td>
<td>India</td>
<td>50</td>
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<tr>
<td>6.</td>
<td>Wyatt et al. (1998)</td>
<td>Mexico</td>
<td>50</td>
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<tr>
<td>7.</td>
<td>Shrestha et al. (2003)</td>
<td>Nepal</td>
<td>50</td>
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<tr>
<td>9.</td>
<td>Tseng et al. (2005)</td>
<td>Taiwan</td>
<td>10</td>
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<tr>
<td>10.</td>
<td>Oremland et al. (2004)</td>
<td>USA</td>
<td>10</td>
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**As exposure and effects on health**

One report of IARC (2012) stated that, in India, Bangladesh, Taiwan, Chile, and the United States about 100 million people worldwide are exposed to As, particularly through ingestion of contaminated food and water exposures also occur through inhalation, especially in agricultural and industrial settings (Nordstrom, 2002).
Human population is mostly exposed to arsenic through ingestion, inhalation and dermal contact. Chen et al. (2004) reported that world population are exposed to arsenic through contaminated water, foods, drugs, wines, smoke of cigarette and fossil fuels. In case of occupational exposure, through industrial exposure, the workers are exposed to airborne arsenic (USPHS, 1989). At the time of treating some disease like syphilis, asthma, rheumatism, cough, pruritus and itching, arsenic contamination has been occurred (Ko, 1999). To treat advanced trypanosomiasis and acute promyelocytic leukemia (APL), pentavalent arsenic is used (Novick and Warrell, 2000). Inorganic As exists in the environment as arsenite (As) or arsenate (As) and is metabolized in humans via conversion of arsenite to arsanate with subsequent methylation to mono- and di-methylated arsenicals (MMA and DMA, respectively) (Drobna, 2005). MMA is considered the most toxic arsenical in vitro (Petrick, 2001) and individuals who excrete a higher proportion of ingested As as urinary MMA have increased risks of As-associated cancers (Steinmau et al, 2006), suggesting a key role for MMA in As toxicity. In addition, inorganic arsenic has been classified as a known human carcinogen (IRIS, 1997). Acute exposures to large doses of inorganic arsenic may be fatal. Several studies showed that, for humans, the lethal dose of orally administered arsenic trioxide has been variously estimated as 70 to 180 mg, 200 to 300 mg (1 to 5 mg/kg for adults weighing 60–75 kg; (Winship, 1984). But, survival has been reported after doses as high as 10 g (Winship, 1984). After lower, repeated doses (e.g., 4 to 15 mg As/d; (Winship, 1984), arsenicals can elicit effects on the nervous, gastrointestinal, and integumentary systems. Chronic exposure to inorganic arsenic can lead to cancer of the skin (hyperpigmentation, hyperkeratosis and hypopigmentation), lungs, bladder and liver (Yeh et al., 1968; Cebrian et al., 1983; Aposhian et al., 2004; IARC, 2012) as well as neurological disorders, muscular weakness, loss of appetite, and nausea. Through acute poisoning of As, oesophageal and abdominal pain, vomiting and diarrhea has occurred (Duker, 2005). There are several proposed mechanisms of toxicity include imbalance between pro-oxidant and antioxidant homeostasis that results in oxidative stress. Oxidative stress occurred due to the inhibition of genetic expression in the cell (Shila et al., 2005; Flora, 2011; IARC, 2004; Tokar, 2011).

**Oxidative stress**

The natural antioxidant defence mechanisms saturate due to the excess production of highly reactive molecules such as reactive oxygen species (ROS) and reactive nitrogen species (RNS) (Maritim, 2003). Most widely studied examples of ROS are superoxide (·O2⁻), hydroxyl ion (·OH) and hydrogen peroxide (H₂O₂) (Evans, 2002). Common RNS include nitric oxide (·NO) and peroxynitrite (ONOO⁻) (Evans, 2002). Oxidative stress induced by oxidant species occurs under conditions when antioxidant defence is attenuated or when the rate constants of the radical reactions are greater than the antioxidant defence mechanisms (Fig. 2; Buettner, 1993).

![Fig. 2: Oxidative stress induced by oxidant species (Source: Buettner, 1993).](image-url)
Source of oxidative molecules

Non enzymatic sources

Fenton's and Haber's reactions: Fridovich (1984) explain the reduction of molecular oxygen to form superoxide anions. These superoxide anions have the ability to form excess and highly reactive oxygen species. The dismutation of superoxides forms hydrogen peroxide ($H_2O_2$).

\[ O^2^- + O^2^- + 2H = H_2O_2 + O_2 \]

Hydrogen peroxide is more stable than $O^2-$ superoxide. It is permeable to plasma membrane and plays two important roles in the body. If it is not scavenged by catalase/GSH (glutathione peroxidase) enzyme defence system then it will occurred highly reactive oxygen species (Halliwell and Gutteridge, 1989). Produced hydrogen peroxides react with metal iron or copper to form more highly reactive hydroxyl ions (OH) through Fenton's reaction.

\[ Fe^{2+} + H_2O_2 = Fe^{3+} + OH + OH \]

Through Haber-Weiss reaction

\[ O_2 + H_2O_2 \rightarrow O_2 + OH + OH \]

$H_2O_2$ reacts with $Cl^-$, $Br^-$, $I^-$ and is utilized by myeloperoxidase to form more reactive hypochloric acid/hyperchloirite. This is important for protein aggregation and fermentation (Babior, 2000).

Enzymatic sources

ROS are generated by oxygen metabolism. They have single unpaired electron in their outer orbit which caused them to become highly reactive. It is produced in all aerobic organisms to perform cellular metabolisms. Xanthine oxidases, cyclo-oxygenases (COX) and lipoxygenases (LOX), NO synthases (nitric oxide synthase) and mitochondrial oxidases are the main enzymes for the source of ROS (Lambeth, 2004). As well as, monoamine oxidase, NADPH oxidase/Respiratory Burst Oxidase (RBO), xanthine oxidoreductase, cytochrome $P_{450}$ oxidase, myelo peroxidase are responsible for the production of oxidative molecules (Cadenas, 2000; El-Benna, 2005; Judge, 2004; Ivanov, 2005; Klebanoff, 2005).

Generation of ROS in cells

Mitochondria, endoplasmic reticulum (ER) and phagocytic cells (neutrophils and other phagocytes) are the major sites for the production of oxidative molecules in the cell (Tatoyan, 1998; Bedard, 2007; Ghosh, 1997).

Other sources

Apoptosis, auto-oxidation of small molecules, peroxisomes and lysosomes are also responsible for the several type of oxidative molecules production (Kam, 2000; Freeman, 1982; Tolbert, 1981; Klebanoff, 2005).

Arsenic and oxidative stress

Oxidative stress is a relatively new theory of arsenic toxicity (Kitchin, 2001). Recent studies have also indicated that arsenic exerts toxicity by generating reactive oxygen species, but the mechanism is still unclear (Chang et al., 2007). Arsenic generates ROS and free radicals like hydrogen peroxide ($H_2O_2$), hydroxyl radicals species ($HO$-), nitric oxide ($NO$), superoxide anion ($O_2^-$), dimethyl arsenic peroxyl radical [(CH$_3$)$_2$AsOO$^-$], and dimethyl arsenic radical [(CH$_3$)$_2$As]$^-$ (Yamanaka et al., 1997; Chen et al., 1998; Gurr et al., 1998; Lynn et al., 2000). Since about 1990, there is the proposal that all of these reactive species are responsible for the stress response elicited by arsenicals. However, the mechanism for the production of this reactive intermediate is still not fully understood. But, Yamanaka et al. (1997) proposed the formation of intermediary arsenic species.

Dimethylarsine, a trivalent arsenic form, is a minor in vivo metabolite of DMA which is a pentavalent arsenic form; produced by a process of reduction in the in vivo condition. This can react with molecular oxygen form a (CH$_3$)$_2$As radicals and superoxide anions. Subsequently, (CH$_3$)$_2$As can add another molecule of molecular oxygen and form the (CH$_3$)$_2$AsO$^-$ radical. Exposure to these free radicals can lead to DNA damage (single strand breaks), (Yamanake and Okada, 1994; Kitchin, 2001). Arsenic also reduces antioxidant levels in plasma, which may accelerate disease development at target site. These reports suggested that intracellular peroxide level is correlated with arsenic induced cellular apoptosis (Hsueh et al., 1998). Ito et al. (1998) reported that GSH has the important protective role against arsenic induced oxidative damage in the cell. Ramos et al. (1995) also described lipid peroxidation as one of the mechanisms of arsenic toxicity in female rats.

Arsenic caused subsequent decrease in cellular GSH concentration. Mainly in the liver, this is inversely

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correlated with lipid peroxidation but not in other tissues (Santra et al., 1999). The functions of glutathione related enzymes like glutathione peroxidase (GPx) and glutathione reductase has either directly or indirectly role as the antioxidant. In case of glutathione-S transferase (GST), it plays the important role in metabolic detoxification. Oxidative stress can be involved in initiation, promotion, or progression of several disorders (Guyton and Kensler, 1993).

In humans, absorbed inorganic arsenic (pentavalent) is biotransformed to trivalent arsenic. Trivalent form of arsenic undergoes methylation to form less toxic compounds. This are excreted in urine but some inorganic arsenic is excreted in the urine unchanged (Hall, 2002). These forms of Arsenic can attenuate various enzymatic action including glycolysis and TCA cycle by binding to sulphhydryl groups of enzymes. Aarsenic compound (pentavalent) has the great role on uncoupling the mitochondrial oxidative phosphorylation. Inhalation exposure to arsenic is associated with an increased risk of lung cancer. The lung is one of the major organs that are affected by arsenic (Kitchin, 2001). Oxidative stress theory for arsenic carcinogenicity can be partially explained by its ability to cause cancer at high rates in the lung, bladder and skin. In case of human lung, it may be an organ responsive to arsenic carcinogenesis because of high partial pressure of oxygen and the fact that dimethylarsine a gas excreted via the lungs (Yamanaka and Okada, 1994). However, oxidative stress has been widely accepted as a general mechanism of action for cellular injury and toxic effects of arsenic (Del Razo et al., 2001).

**Conclusion**

Arsenic exposure can affect millions of people worldwide. Through different types of exposure it exposed to animals and caused severe types of different disorders in the animal body. Arsenic can affect the mammalian body by inducing the oxidative stress and cause of different diseases. Arsenic is a main agent of misbalancing between the generation of oxidative molecules and antioxidant defense system. Ultimately, it caused cancer in different organs of the body and as well as cause of cardiovascular disease, skin lesions, urinary disorder, metabolic disorders, etc. However, we are aware about the arsenic as a causal agent of oxidative stress but its proper mechanism yet to be known. To provide a deeper understanding of the pathology of arsenic induced oxidative stress diseases and the toxicology of arsenic in various organs, further research is necessary.

**Conflict of interest statement**

Authors declare that they have no conflict of interest.

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